

## Soil Erodibility in European Mountain Beech Forests

Journal:	<i>Canadian Journal of Forest Research</i>
Manuscript ID	cjfr-2020-0361.R2
Manuscript Type:	Article
Date Submitted by the Author:	17-Aug-2021
Complete List of Authors:	<p>Kasanin-Grubin, Milica; University of Belgrade, Institute of Chemistry, Technology and Metallurgy          Hukić, Emira ; University of Sarajevo, Faculty of Forestry          Bellan, Michal; Mendel University in Brno, Department of Forest Ecology          Bielak, Kamil; Warsaw University of Life Sciences, Department of Silviculture          Bosela, Michal; Faculty of Forestry Technical University in Zvolen, ; National Forest Centre, Forest Research Institute Zvolen,          Coll, Lluís; University of Lleida, Department of Agriculture and Forest Engineering (EAGROF); Joint Research Unit CTFC - AGROTECNIO, 7Joint Research Unit CTFC - AGROTECNIO          Czacharowski, Marcin; Warsaw University of Life Sciences, Institute of Forest Sciences, Department of Silviculture          Gajica, Gordana; University of Belgrade, Institute of Chemistry, Technology and Metallurgy, National Institute of the Republic of Serbia          Giammarchi, Francesco; Libera Università di Bolzano, Facoltà di Scienze e Tecnologie          Gömörková, Erika ; Technical University of Zvolen          del Rio, Miren; CIFOR-INIA,          Dinca, Lucian; National Institute for Research and Development in Forestry Marin Dracea, National Institute for Research and Development in Forestry Marin Dracea          Djogo Mračević, Svetlana; University of Belgrade, Faculty of Pharmacy          Klopčič, Matija; University of Ljubljana, Biotechnical Faculty, Department of Forestry and Renewable Forest Resources          Mitrović, Suzana; Institute of Forestry          Pach, Maciej; University of Agriculture in Krakow, Department of Forest Ecology and Silviculture          Randjelović, Dragana; Institute for Technology of Nuclear and Other Mineral Raw Materials          Ruiz-Peinado, Ricardo; Spanish Institute for Agriculture and Food Research and Technology (INIA), Forest Research Center (CIFOR)          Skrzyszewski, Jerzy ; University of Agriculture in Krakow, Department of Forest Ecology and Silviculture          Orlić, Jovana; University of Belgrade Faculty of Chemistry          Štrbac, Snežana; University of Belgrade, Institute of Chemistry, Technology and Metallurgy, National Institute of the Republic of Serbia          Stojadinović, Sanja; University of Belgrade, University of Belgrade, Institute of Chemistry, Technology and Metallurgy, National Institute of the Republic of Serbia</p>

	<p>Tonon, Giustino; Libera Università di Bolzano Facoltà di Scienze e Tecnologie, Facoltà di Scienze e Tecnologie, Libera Università di Bolzano</p> <p>Tosti, Tomislav; University of Belgrade</p> <p>Uhl, Enno; Technical University of Munich, Forest Growth and Yield Science, School of Life Sciences, Weiherstephan; Bavarian State Institute of Forestry (LWF)</p> <p>Veselinović, Gorica; University of Belgrade, , Institute of Chemistry, Technology and Metallurgy, National Institute of the Republic of Serbia</p> <p>Veselinović, Milorad; Institute of Forestry</p> <p>Zlatanov, Tzvetan; Bulgarian Academy of Sciences, Institute of Biodiversity and Ecosystem Research</p> <p>Tognetti, Roberto; Università degli Studi del Molise, Dipartimento di Agricoltura, Ambiente e Alimenti; Università degli Studi del Molise, Centro di Ricerca per le Aree Interne e gli Appennini (ArIA); EFI Project Center on Mountain Forests</p>
Keyword:	bedrock, environmental change, geochemistry, soil organic matter, CLIMO Cost Action
Is the invited manuscript for consideration in a Special Issue? :	CLIMO 2019

SCHOLARONE™  
Manuscripts

## 1 Soil Erodibility in European Mountain Beech

2 Milica Kašanin-Grubin<sup>1</sup>, Emira Hukić<sup>2</sup>, Michal Bellan<sup>3</sup>, Kamil Bialek<sup>4</sup>, Michal Bosela<sup>5</sup>, Lluís Coll<sup>6,7</sup>, Marcin  
3 Czacharowski<sup>4</sup>, Gordana Gajica<sup>1</sup>, Francesco Giammarchi<sup>8</sup>, Erika Gömöryová<sup>5</sup>, Miren del Río<sup>9,10</sup>, Lucian  
4 Dinca<sup>11</sup>, Svetlana Đogo Mračević<sup>12</sup>, Matija Klopčič<sup>13</sup>, Suzana Mitrović<sup>14</sup>, Maciej Pach<sup>15</sup>, Dragana Randjelović<sup>16</sup>,  
5 Ricardo Ruiz-Peinado<sup>9,10</sup>, Jerzy Skrzyszewski<sup>15</sup>, Jovana Orlić<sup>17</sup>, Snežana Štrbac<sup>1</sup>, Sanja Stojadinović<sup>1</sup>,  
6 Giustino Tonon<sup>8</sup>, Tomislav Tosti<sup>17</sup>, Enno Uhl<sup>18,19</sup>, Gorica Veselinović<sup>1</sup>, Milorad Veselinović<sup>14</sup>, Tzvetan  
7 Zlatanov<sup>20</sup>, Roberto Tognetti<sup>21,22,23</sup>

8 <sup>1</sup> University of Belgrade, Institute of Chemistry, Technology and Metallurgy, National Institute of the Republic of  
9 Serbia (Serbia) mkasaningrubin@chem.bg.ac.rs, gordana.gajica@ihtm.bg.ac.rs,

10 snezana.strbac@ihtm.bg.ac.rs, sanja.stojadinovic@ihtm.bg.ac.rs, goricagrbovic@chem.bg.ac.rs

11 <sup>2</sup> University of Sarajevo, Faculty of Forestry (Bosnia and Herzegovina) e.hukic@sfsa.unsa.ba

12 <sup>3</sup> Mendel University in Brno, Department of Forest Ecology (Czech Republic) bellan.m@seznam.cz

13 <sup>4</sup> Warsaw University of Life Sciences, Institute of Forest Sciences, Department of Silviculture

14 (Poland) kamil.bielak@wl.sggw.pl, marcin.czacharowski@wl.sggw.pl

15 <sup>5</sup> Technical University in Zvolen, Faculty of Forestry, (Slovakia) ybosela@tuzvo.sk, gomoryova@tuzvo.sk

16 <sup>6</sup> University of Lleida, Department of Agriculture and Forest Engineering (EAGROF), (Spain) lluis.coll@udl.cat

17 <sup>7</sup> Joint Research Unit CTFC - AGROTECNIO (Spain) lluis.coll@udl.cat

18 <sup>8</sup> Libera Università di Bolzano, Facoltà di Scienze e Tecnologie, (Italy) giustino.tonon@unibz.it,

19 francesco.giammarchi@unibz.it

20 <sup>9</sup> Forest Research Center, INIA, CSIC National Institute for Agricultural and Food Research and Technology

21 (INIA), Forest Research Center (CIFOR), (Spain) delrio@inia.es, ruizpein@inia.es

22 <sup>10</sup> Sustainable Forestry Management Research Institute (iuFOR), UVa-INIA (Spain) delrio@inia.es,

23 ruizpein@inia.es

24 <sup>11</sup> National Institute for Research and Development in Forestry 'Marin Dracea', (Romania)

25 dinka.lucian@gmail.com

26 <sup>12</sup> University of Belgrade, Faculty of Pharmacy (Serbia) svetlana.djogo@pharmacy.bg.ac.rs

27 <sup>13</sup> University of Ljubljana, Biotechnical Faculty, Department of Forestry and Renewable Forest Resources

- 28 (Slovenia) matija.klopcic@bf.uni-lj.si
- 29 <sup>14</sup> Institute of Forestry (Serbia) mitrovicsuzana79@gmail.com, mvčetiri@gmail.com
- 30 <sup>15</sup> University of Agriculture in Krakow, Department of Forest Ecology and Silviculture, (Poland) rlpach@cyf-  
31 kr.edu.pl, rlskrzys@cyf-kr.edu.pl
- 32 <sup>16</sup> Institute for Technology of Nuclear and Other Mineral Raw Materials (Serbia) d.randjelovic@itnms.ac.rs
- 33 <sup>17</sup> University of Belgrade, Faculty of Chemistry (Serbia) jovanaorlic@chem.bg.ac.rs, tosti@chem.bg.ac.rs
- 34 <sup>18</sup> Technical University of Munich, School of Life Sciences Weihenstephan, Chair of Forest Growth and Yield  
35 Science, (Germany) enno.uhl@tum.de
- 36 <sup>19</sup> Bavarian State Institute of Forestry (LWF) (Germany) (enno.uhl@tum.de)
- 37 <sup>20</sup> Institute of Biodiversity and Ecosystem Research, Bulgarian Academy of Sciences (Bulgaria)  
38 tmzlatanov@gmail.com
- 39 <sup>21</sup> Università degli Studi del Molise, Dipartimento di Agricoltura, Ambiente e Alimenti (Italy) tognetti@unimol.it
- 40 <sup>22</sup> Università degli Studi del Molise, Centro di Ricerca per le Aree Interne e gli Appennini (ArIA), (Italy)  
41 tognetti@unimol.it
- 42 <sup>23</sup> The EFI Project Centre on Mountain Forests (MOUNTFOR) (Italy) tognetti@unimol.it

#### 43 Abstract

44 Forests in Europe are, at present not endangered by soil erosion, however, this can change with climate  
45 change or intensified forest management practices. Using a newly established network of plots in beech  
46 forests across Europe, the aims of this study were 1) discrimination of soil properties and erodibility indices in  
47 relation to bedrock, 2) determination of geochemical properties and Corg influencing erodibility, and 3)  
48 assessment of the effect of soil depth on erodibility indices. Seventy-six soil samples from 20 beech forests  
49 were collected in 11 countries to quantify soil properties influencing erodibility indices clay ratio, modified clay  
50 ratio, sodium adsorption ratio, and oxides ratio. Results indicate that dominant soil properties, determined by  
51 bedrock, that correlate with forest soil  
52 erodibility indices are: Corg, pH, EC, Ca and Na ion concentrations, total-water soluble cations, and the % of  
53 sand. According to the tested indices, soil susceptibility to erosion follows the sequence:  
54 granite>andesite>sandstone>quartzite>limestone. Deeper soil horizons on granite are more susceptible to

55 erosion than surface horizons, while this is not the case for soils on limestones. In conclusion, forest  
56 management should consider the predisposition of different soil types to erosion.

57 Keywords: bedrock, environmental change, soil erodibility, texture, organic  
58 matter, geochemistry, CLIMO Cost Action

## 59 Introduction

60 At the forest site-level erodibility depends on physical and chemical properties of soil that are mainly a function  
61 of the bedrock material (Milodowski et al. 2015 a, 2015b). It is a measure of general susceptibility to the  
62 detachment and transport of soil particles by erosion processes and forces, varying spatially and by soil depth.  
63 However, the impact of bedrock on soil degradation through erosion is not sufficiently understood (Jiang et al.  
64 2020) and changing environmental conditions (Christensen et al. 2013, Hartmann et al. 2013) influencing  
65 geochemical processes can alter soil properties. It is known that soil physical and chemical features determine  
66 soil erodibility, but it is less well known how their interactions alter erodibility (Wang et al. 2013). Considering  
67 the heterogeneous nature of soils more information is needed for a better understanding of the effect of  
68 primary particles on site-specific soil erodibility.

69 Indices of erodibility are based on several key properties interacting with each other. Some of the most  
70 important to be included in assessments of soil erosion resistance are particle size distribution and  
71 geochemical properties like clay mineralogy, sodium adsorption, relative cation content, pH and organic C  
72 ( $C_{org}$ ) content (Grabowski et al. 2010). These also represent the most commonly analyzed soil properties,  
73 known from forest ecosystem surveys and monitoring (ICP Forest Program, [www.icp-forests.net/page/level-ii](http://www.icp-forests.net/page/level-ii)).  
74 Relative proportions of different-sized particles substantially affect erodibility. Sandy and silty soils, due to their  
75 uncohesive nature, have a small inherent resistance to erosion (Parlak 2009). Many clayey soils are sensitive  
76 to surface runoff and their erodibility increases when the clay content is greater than 50%, in contrast with an  
77 increase in clay content up to 30-50%, which improves particles' resistance to erosion (Grabowski et al. 2010).  
78 The erodibility of clay soils is related to clay mineralogy and adsorbed cations. Erodibility is closely connected  
79 with the dispersion potential of clay minerals normally measured by sodium adsorption ratio (SAR)  
80 (Rengasamy et al. 1984). Clay erodibility also has been examined through the oxides ratio (Bennett 1926).  
81 Some authors use quantitative relationships of soil particle size distribution,  $C_{org}$ , soil permeability, and soil

82 structure (Wang et al. 2016).

83 The presence of stabilizing substances, like organic matter, oxides, carbonates and cations in soil solution,  
84 depends on the soil depth. The resistance to erosion most commonly decreases with the increase in soil depth  
85 (Bouyoucos 1935). In soils rich in  $C_{org}$ , particle size density can be a good erodibility indicator, while in humus-  
86 poor soils lack of humus ( $C_{org}$  lower than 2%) or quantitative relation of particle size distribution and cation ratio  
87 can be better indicators (Liu and Han 2020). Since soil characteristics vary by depth, different horizons might  
88 express different erodibility properties. By analyzing vertical variability of soil erodibility traits one can assess  
89 the entire soil's resistance to erosion.

90 Under forest land use soil erosion is currently of minor importance in Europe (Borelli et al. 2006). Although  
91 forest soils are often found on steep terrain, a well-established root system and closed canopy cover prevents  
92 those soils from eroding. This is an argument for applying forestry management practice that maintains forests  
93 with a closed canopy. However, the resistance to erosion of forest soils can dramatically change in the wake of  
94 climate warming, with increasing occurrence of disturbance events such as outbreaks of pests and pathogens,  
95 for instance, satellite imagery reveals that the average patch size of harvested area, in reference to the period  
96 from 2004 to 2018, has increased by 34 percent across Europe, with a potential effect on soil erosion  
97 (Ceccherini et al. 2020). Although there are limitations in using erodibility indices, since each is specific to a  
98 certain erosion wildlife populations, or fire and concerning changes in forest management (Haas et al. 2020).  
99 process and force, they could be useful for monitoring changes in soil resistance to erosion in a changing  
100 environment, and for planning soil protective silvicultural measures.

101 European beech (*Fagus sylvatica* L.) is naturally distributed throughout much of Europe (EUFORGEN 2009)  
102 on a wide range of geological substrates (Leuschner et al. 1996). Because it is one of the most represented  
103 species throughout Europe, beech stands are ideal for determining criteria that would allow assessment of the  
104 forest soil/bedrock sensitivity to erosion. This study is part of the European Cooperation in Science and  
105 Technology (COST) action CA15226 "Climate Smart Forestry in Mountain Regions" known as CLIMO  
106 ([www.cost.eu/actions/CA15226](http://www.cost.eu/actions/CA15226)). The CLIMO program is an integrative approach that aims to identify site-  
107 specific management practices to promote adaptive forest management including measures for soil protection  
108 (Bowditch et al. 2020). This study seeks possible erodibility indicators applicable for forest monitoring, which

109 can be calculated from commonly monitored features under the forest surveys.

110 On 20 plots from the CLIMO COST Action network we set up the study focused on 1) discrimination of soil  
111 properties and erodibility indices in relation to bedrock material, 2) determination of geochemical and  $C_{org}$   
112 properties influencing soil erodibility, and 3) assessing the effect of soil depth on geochemical properties and  
113 erodibility indices. With these objectives, we analyzed four erodibility indices: clay ratio (CR) (Bouyoucos,  
114 1935), modified clay ratio (MCR) which considers soil  $C_{org}$  (Kusre et al., 2018), sodium adsorption ratio (SAR)  
115 (Rengasamy et al., 1984) and oxides ratio (Bennet, 1926), chosen based on the most widely assessed soil  
116 properties. The results obtained from 20 pure beech forest stands from 11 European countries are contributing  
117 to the knowledge on climate-smart forestry by showing how soils differ in their susceptibility to erosion and how,  
118 among other things, erodibility should be considered when planning forest management measures.

## 120 Methods

### 121 Study sites and field sampling

122 From a total of 70 plots established within the CLIMO COST Action network, we selected 20 with predominant  
123 bedrock and soil types in Europe to assess inherent soil resistance to erosion pressure in beech forests  
124 (Figure 1, Table 1). Selected plots corresponded to five types of bedrock: granite (G; 5 plots), andesite (A; 1  
125 plot), sandstone (S; 5 plots), quartzite (Q; 1 plot) and limestone and dolomite (L; 8 plots), , , , and (Table 1).  
126 Altitudes of plots vary between 415 and 1461 m a.s.l. and the slope ranges between 2.7 and 32 degrees. The  
127 mean annual air temperature (MAT) in these locations based on the observed period between 1961 and 1990  
128 ranges from 3.4 to 10.5 °C, and mean annual precipitation (MAP) from 520 to 1100 mm. We included fully  
129 stocked unmanaged or just slightly managed (no silvicultural treatment over the last 10 years) stands,  
130 representing the natural dynamics of mountain beech forests and climate variation across Europe.

131 For this study, 76 soil samples from 20 plots in 11 countries (Bosnia and Herzegovina-BA, Bulgaria-BG, Czech  
132 Republic-CZ, Germany-GE, Italy-IT, Poland-PL, Romania-RO, Serbia-SRB, Slovakia-SK, Slovenia-SL, and  
133 Spain-SP) were collected during the fall of 2018 (Figure 1). One representative soil profile was analyzed per

134 plot and general site characteristics are given in Table 1. Soil samples were collected at four depths in the  
135 profile 0-10 cm, 10-20 cm, 20-40 cm, and 40-80 cm, except at site 1 (BA1: soil depth was only 20 cm), 4 and 5  
136 (CZ-1 and CZ-2: soil depth was only 40 cm). Approximately one kilogram of soil sample was taken from across  
137 a 1 m wide soil horizon, following the procedure given in the ICP Forest manual (Cools and De Vos 2016).  
138 Samples were labeled, stored in plastic bags, and shipped to the University of Belgrade where laboratory  
139 analyses were carried out.

#### 140 Laboratory work

141 The Manual for sampling and analysis of soils (Cools and De Vos 2016) was followed for the selection and  
142 procedure of soil analyses. The following characteristics were determined on all samples: particle size  
143 distribution, the content of organic carbon ( $C_{org}$  in %), pH values, electrical conductivity (EC in  $\mu\text{S}$ ), the content  
144 of  $\text{CaCO}_3$  (%), nitrogen (N), C/N ratio, the concentration of water-soluble cations  $\text{Al}^{3+}$ ,  $\text{Ca}^{2+}$ ,  $\text{Mg}^{2+}$ ,  $\text{K}^+$ ,  $\text{Na}^+$ ,  
145  $\text{Mn}^{2+}$ ,  $\text{Fe}^{3+}$ , total concentrations of elements Na, Mg, Al, Si, K and Ca (%) and content of  $\text{SiO}_2$ ,  $\text{Al}_2\text{O}_3$ , and  
146  $\text{Fe}_2\text{O}_3$  oxides.

#### 147 Particle size analyses

148 Grain size analysis was performed according to a standard wet sieving procedure (Dane and Topp 2002) using  
149 a set of sieve sizes ranging from 2.0 to 0.063 mm (2, 1, 0.5, 0.25, 0.125 and 0.063 mm). The sieved material  
150 was dried in an oven at 105 °C and weighed. A standard sedimentation procedure (pipette analysis) was  
151 performed for <0.063 mm fractionation of particles. Sodium hexametaphosphate (3.3 %) and an ultrasound  
152 bath were used as dispersing agents.

#### 153 Measurements of pH values and electrical conductivity (EC)

154 Soil samples were analyzed for pH and EC in water using a 1:5 soil/water suspension. For these studies, a  
155 suspension of 4 g of soil and 20 ml of distilled water was used. The prepared sample was dispersed in an  
156 ultrasonic bath for 3 minutes, then the soil solutions were centrifuged at 10000 rpm for 20 minutes and filtered  
157 through a cellulose filter with 1 micron pore size. The pH values were determined using pH meter WTW  
158 INOLAB pH 720 (Welheim, Germany) equipped with a glass electrode. The conductivity measurements were



159 performed using WTW INOLAB 7110 conductometer (Welheim, Germany).

160 Content of organic carbon ( $C_{org}$ ) and nitrogen (N)

161 Soil samples were pulverized to a fine powder. Subsequently, the samples were placed in the oven at the  
162 temperature of 105 °C to eliminate hygroscopic moisture. Afterward, the dry residue was pre-treated with  
163 diluted hydrochloric acid (1:3, v:v) to eliminate carbonates. After the carbonates were removed, elemental  
164 analysis was performed to determine the contents of organic carbon. The measurements were done using a  
165 Vario EL III, CHNS/O Elemental Analyzer (Elementar Analysensystem GmbH, Germany).

166 Concentrations of major elements

167 Concentrations of major elements (Na, Mg, Al, Si, K, Ca, Fe) were determined using the X-ray fluorescence  
168 (XRF) method. After drying until constant mass at 105 °C, samples were prepared as pressed pellets by mixing  
169 soil and tableting aid wax (Hoechst wax micro powder produced by Merck, Lot number-K36429014636), at a  
170 ratio of 85:15, respectively. The pressure of 25t in a Retsch PP25 hydraulic press was applied for 5 min to the  
171 mixture to produce stable pellets which were 32 mm in diameter and approximately 3 mm thick. A Spectro Xepos  
172 Energy Dispersive X-ray Fluorescence spectrometer (XRF, Germany), equipped with a binary cobalt/palladium  
173 alloy thick-target anode X-ray tube (50W/60kV) and combined polarized/direct excitation was used. The analysis  
174 was performed with a high-resolution silicon drift detector (SDD) with an air-cooling system. For qualitative  
175 analysis, spectral recording, and data processing, a software program Xepos C and Fundamental Parameters  
176 (JRRM) method was used.

177 Concentrations of water-soluble cations

178 Concentrations of water-soluble cations  $Al^{3+}$ ,  $Fe^{3+}$ ,  $Mn^{2+}$ ,  $Ca^{2+}$ ,  $Na^{+}$ ,  $K^{+}$ ,  $Mg^{2+}$  were determined using the 1:10  
179 soil/water dilution, following the procedure given by Faulkner et al. (2001). Total water-soluble cations (TWSC)  
180 were calculated as the sum of all cations.

181 Erodibility indices

182 To assess the erodibility of soils, four of the most common indices were tested: clay ratio (CR; Bouyoucos

183 1935), modified clay ratio (MCR; Kusre et al. 2018), sodium adsorption ratio (SAR; Rengasamy et al. 1984)  
184 and the oxides ratio of  $\sum \text{SiO}_2 / (\text{Al}_2\text{O}_3 + \text{Fe}_2\text{O}_3)$  (Bennett 1926). Bouyoucos (1935) proposed the CR, ratio  
185 between sand+silt (%) / clay (%), as a measure of binding due to the presence of clay and it is inversely  
186 related to soil erodibility, because high specific surface area clay particles are more reactive than coarse  
187 particles and can store higher amounts of carbon than sandy soils (Sulman et al. 2014). The MCR, ratio  
188 between sand+silt / clay+soil organic matter (%), might be more suitable for soils having high organic content  
189 (Kusre et al., 2018), which is true for most forest soils. The SAR was calculated using the following equation:  
190  $\text{SAR} = [\text{Na}] / ([\text{Ca} + \text{Mg}] / 2)^{0.5}$  where all concentrations are in mmol/L.

191 The applicability of four erodibility indices varies among different soil textural categories and also their  
192 referential values (Table 2). As mentioned previously, SAR and oxides ratio represents the means to assess  
193 erodibility of clayey soil, while CR and MCR values are suitable for all soil textural categories. Considering the  
194 stable physical characteristics of soils, SAR values are employed to try to predict the response of soils to  
195 environmental change.

## 196 Statistics

197 The similarity/dissimilarity of soils was assessed using Principal Component Analysis (PCA) including all  
198 analyzed soil features and soil erodibility indices ( $N=75$ ). To test differences in erodibility indices between the  
199 bedrock types, a one-way analysis of variance (ANOVA) was performed. All values were square-root  
200 transformed for achieving normal distributions and equal variances. Pearson correlation was used to quantify  
201 relationships among soil erodibility indicators and selected physical and chemical properties. The variables  
202 were again square-root transformed. Statistical analyses were performed by using Minitab 19 (Minitab, LLC,  
203 2021).

## 204 Results

### 205 Soil properties relative to bedrock

206 The soils used in this study were developed on 5 bedrock types: granite (G), andesite (A), sandstone (S),  
207 quartzite (Q), and limestone & dolomite (L), with marked differences in texture and physio-chemical properties.

208 Principal component analysis discriminated soils according to these five bedrocks groups with eigenvalues  
209 (Figure 2), describing how the variables are spread, 6.91 (PCA1), 5.43 (PCA2), and 2.49 (PCA3). Furthermore,  
210 PCA confirmed that the differentiation of soil characteristics is primarily determined by pH, EC,  $C_{org}$ , total and  
211 water-soluble contents of Ca and Mg in soils on limestones, the contents of Si, Al, K, Na for soils on  
212 sandstone, and the content of sand for soils on granites. Although erodibility indices will be discussed later, this  
213 result indicates that these parameters should be key parameters for determining if the soil is prone to erosion.

214 Soil texture varied considerably among bedrock groups (Figure 3). Soils on G and A bedrock were  
215 characterized as sandy clay to sandy clay loams. Soils on S and Q were sandy clays and on L mainly clays to  
216 sandy clays. On G, the soils had the lowest clay content (mean of 41.3 %) with the highest content of sand  
217 (47.5 %), soils on S, A, and Q had medium clay content (65.6 %) and soils on L showed the highest clay  
218 content (71.3 %) among the groups. These fractions were used further to describe erodibility indices.

219 Chemical soil properties also varied among soils grouped based on bedrock (S1; S2). Total amounts of Na,  
220 Mg, Al, Fe, K, Ca were similar among the groups. The most significant differences between analyzed groups  
221 were found for the amount of total Si, for which the highest values were in the soils on bedrock group S, and of  
222 Ca and Mg, with the highest values in soils in group L.

223 The mean pH values of 4.81 show that soils on G are more acidic than on L with a pH of 6.04. For all analyzed  
224 soil profiles, pH increased with an increase of a soil depth indicating cation depletion processes. Electrical  
225 conductivity for most soils pointed to a high ionic soil activity. Values of EC varied between 47.1  $\mu\text{S}$  and 74.2  
226  $\mu\text{S}$  on A, G, Q, and S, while a higher value of 112  $\mu\text{S}$  was determined on L.

227 As an indication of the soil resistance to depletion and dispersion, values of TWSC varied between 99.7  $\mu\text{g/g}$   
228 (A) and 313.2  $\mu\text{g/g}$  (L). The lowest average concentrations of Ca ion were found in G and S (22.5  $\mu\text{g/g}$  and  
229 25.2  $\mu\text{g/g}$ , respectively) and the highest was found in L (188.6  $\mu\text{g/g}$ ). We found the highest amount of Mg ion  
230 also on L (3.1  $\mu\text{g/g}$ ), whereas the lowest was found on Q and A. Concentration of Al ion, an indication of soil  
231 acidification, was highest with 14.5  $\mu\text{g/g}$  in soil on G, followed by S and A. The average amount of  $C_{org}$  varied  
232 between 1.27 (Q) and 4.19 % (L). The largest  $C_{org}$  concentrations (3.2 to 12.4 %) were found in the upper soil  
233 layers (0-10 cm) for Q and L bedrock and the lowest amounts ranged from 0.1 to 0.9 % in deep soil layers (>20

234 cm).

235 Soil erodibility indices concerning bedrock and soil depth

236 With the one-way ANOVA test, we found significant differences ( $p < 0.01$ ) in all four soil erodibility indices CR,  
237 MCR, SAR, and oxides ratio among bedrock groups (Figure 4, S2). Soil erodibility indices decreased in the  
238 following sequence  $G > A > S > Q > L$ . The only deviation from this result was found for the oxides ratio is observed  
239 in S, which showed higher resistance to erosion than soil on L. An increase in soil erodibility indices was  
240 associated with an increase in sand content and a decrease in  $C_{org}$ , TWSC and pH values. Also, soil erodibility  
241 indices CR, MCR, and SAR increased with soil depth for G, A, and Q, while similar erodibility indices were  
242 found across depths for S and L.

243 As shown in Figure 4, the indices of erodibility varied considerably within each bedrock group. Such  
244 differences in erodibility are the result of the variability of soil properties. A high amount of sand and low pH  
245 values, which are associated with the G group, were significantly correlated with the values of CR, MCR, and  
246 SAR (Table 3). The S bedrock group showed contrasting indices when SAR and oxides ratio were compared  
247 to L, which is probably linked to a higher heterogeneity of this group. This indicates the need to combine all  
248 four indices for assessing erodibility of soils on S. A high amount of the  $C_{org}$  indicates the need to include MCR  
249 instead of CR, particularly in topsoil. Soils on L were attributed to the high content of  $C_{org}$  indicating that MCR is  
250 a better indicator of erodibility than CR and SAR. Herein, the value of MCR should be combined with TWSC  
251 and Ca and Mg cation concentrations which were significantly negatively correlated.

252 As shown in Table 3, a significant negative correlation was found between soil  $C_{org}$  and SAR, and between pH,  
253 EC, Ca ions, TWSC and CR, MCR, SAR. A significant positive correlation was found between sand content  
254 and SAR, and also among indicators MCR, SAR, and oxides ratio.

255 All soils show an increasing trend of soil erodibility with an increase in soil depth (Figure 4, S2) with a strong  
256 increasing trend in soil erodibility is observed in G, A and Q groups. The other two groups (S and L) expressed  
257 less vertical dissimilarity in erodibility indices CR, MCR, and SAR. Soil  $C_{org}$  and TWSC in most soils were  
258 decreasing with increasing soil depth, while sand content and pH values increased with soil depth.

259 Discussion

260 Soil erodibility indices in relation to soil and bedrock material

261 Soils on different bedrock expressed significantly different values of erodibility indices. Our results indicate that  
262 tested beech forest soils in the current state are stable and not prone to erosion. Based on susceptibility to  
263 erosion soils followed the sequence  $G>A>S>Q>L$ , where accordingly, soil on G points to highest and soils on L  
264 to lowest susceptibility to erosion.

265 Although our results indicate that these forest soils in their current state are stable and not prone to erosion, it  
266 is expected that they would respond differently under changed environmental conditions. The expected impact  
267 of climate change on perturbations in most forest ecosystems includes more violent weather phenomena,  
268 drought and changes in drying and rewetting cycles (Reichstein et al. 2013). Such weather extremes increase  
269 the release of nutrients due to physical disruption of the soil structure (Bünemann et al. 2013) and loss of soil  
270 organic matter. The highest susceptibility to erosion was found for the soils of a granitic group of bedrock  
271 based on all four observed indices (Figure 4). This is explained by the high amount of non-cohesive fraction in  
272 sandy clays to sandy clay loams. The granitic group of rocks produces a lot of coarse texture regolith due to  
273 weathering processes. It is known that bedrocks that produce more loose and coarse-textured soils are more  
274 prone to erosion (Grabowski et al. 2011). The link between bedrock and texture affecting soil resistance to  
275 erosion in temperate regions is most evident in young soils, i.e. early phases of soil development when they  
276 are most similar to bedrock. This is common for forest soils on hillslopes. Although, granit group expressed the  
277 highest erodibility, with a considerable amount of clay (>40%; Figure 3 and S1) these soils most probably have  
278 the maximum erosion thresholds (Grabowski et al., 2010), due to the influence of clay on hydrodynamic  
279 smoothing, clay/sand adhesion and clay cohesion.

280 Similar erodibility values of CR and MCR are found for andesites, sandstones, quartzite, and limestone-  
281 dolomite groups of bedrock. Erodibility of soils that contain higher percentages of the clay fraction is generally  
282 lower; for better insight, clay chemical and mineralogical properties are important to consider. The high  
283 adsorption capacity of clays may show higher erodibility and, in that case, TWSC may influence higher  
284 susceptibility to erosion. In contrast, cation exchange capacity can also be negatively correlated with soil  
285 erodibility depending on soil water chemistry (Gerbersdorf et al. 2007). Therefore, SAR and oxides ratio are  
286 here important predictors of dispersiveness and detachment. Soils formed on limestone have different

287 properties from the bedrock as defined by the quantity and mineralogy of clay fraction. The clay fraction will  
288 express different levels of erodibility in relation to mineralogy. Sandstones show high variability which is most  
289 likely an effect of different sediment origins and physio-chemical properties. Their erodibility indices should also  
290 be determined using a combination of both soil texture and chemical properties. Such differences initially will  
291 affect variation in erodibility indices. Sandstones that have a high clay content (transitioning to claystone)  
292 express the highest resistance to erosion.

293 Bedrock types chosen in this study are among the most frequent in the lithosphere: granite and granodiorites  
294 account for 22%, sandstones for 1.7, and carbonate rocks for 2% (Amelung et al. 2018). The estimated  
295 erodibility indices can be useful for monitoring forest soils if these values are compared to commonly  
296 determined levels of visual change in the landscape. The range of erodibility values can be determined for  
297 different types of beech forests to be monitored under regular monitoring programs which could be more  
298 sensitive for the prediction of erosion.

#### 299 Influence of soil physical and geochemical properties on soil erodibility

300 In our study, we found that soil erodibility of beech forest soils, determined through erodibility indices, was  
301 significantly influenced by the amount of sand, water-soluble cations, the concentration of Ca ion, the content  
302 of soil  $C_{org}$ , values of pH and EC, which agrees with previous studies (e.g. Grabowski et al. 2011). Regarding  
303 interactions among soil physical and chemical properties, different geological substrates act differently.

304 Properties determining the major character of soil should be considered when estimating erodibility. The major  
305 characteristics describing the granitic group is sand content (Figure 2), which is significantly correlated to CR,  
306 MCR, and SAR. In sandy soils with small amounts of organic matter, the amount of sand is an important  
307 predictor of erosion processes. Soils formed on sandstone were the most heterogeneous group in texture.  
308 Sandstones were strongly characterized by clay content and Fe, Al and Mn cation concentrations (Figure 2).  
309 Therefore, due to a high clay content, erodibility indices hereafter should be combined with oxides ratio values,  
310 SAR and Ca/Mg ratio, to evaluate soils of a similar texture. The soils on limestone and dolomites were  
311 characterized by high Ca and Mg ions concentrations,  $C_{org}$ , pH, and EC values, which are all important to  
312 consider when assessing the potential for soil loss.

313 When looking at CR and MCR, the relationship between texture and erodibility is not straightforward. Although  
314 erodibility is very sensitive to small changes in soil grain size distribution, it should be considered concerning  
315 other physical and chemical soil properties (Wischmeier and Mannering 1969). Two tested erosion indices, CR  
316 and MCR, are primarily based on soil grain size characteristics as a measure of binding due to the presence of  
317 clay, which is inversely related to soil erodibility. Due to a high specific surface area, clay particles are more  
318 reactive than coarse particles and can store higher amounts of carbon than sandy soils (Sulman et al. 2014).  
319 Brayn (1968) indicates that the higher the percentage of clay the more reliable the index is, with a clay  
320 content higher than 10% CR considered reliable. Because this ratio neglects the presence of organic matter,  
321 which is probably even a more important aggregate-cementing agent (Robinson and Page 1950), it should be  
322 used for layers with a low content of organic C. The MCR is more suitable for soils having high organic content  
323 (Kusre et al. 2018) which is true for forest soils. The obtained values of MCR for most forest soils are also  
324 indicating that tested soils are stable (Figure 4).

325 Because the tested indices based on physical characteristics indicated that beech forest soils are generally  
326 more stable, the SAR based on chemical compositions was used to predict the response of soils to  
327 environmental change. It should be underlined that SAR is primarily designed for sodic material, and none of  
328 the forest soils strictly fall into this category. However, we chose SAR as one of the few erosion indicators  
329 based on chemical composition. Sodium adsorption ratio values below 1 indicate stable soils (S2). Rengasamy  
330 et al. (1984) used the SAR/EC ratio to determine the dispersivity status of red-brown soils in Australia. Naidu et  
331 al. (1995) and later Faulkner et al. (2000, 2003) supported the use of this ratio for soils potentially exposed to  
332 erosion. According to the classification given by Rengasamy et al. (1984), which is based on the ratio between  
333 SAR and sum of all cation concentrations (TWSC), the soils in our study are potentially dispersive since SAR is  
334  $< 3$  and the sum of water-soluble cation concentration is  $< 3.8$ .

335 Furthermore, Faulkner et al. (2000) suggested the use of SAR/pH as a site signature due to the buffering role  
336 of calcium. This ratio shows that SAR depends on soil pH values with a correlation of  $-0.633$ . However, two  
337 trends are present. In conditions of  $\text{pH} < 5.6$  the SAR has values ranging from 0.026 to 0.353, while in soils with  
338  $\text{pH} > 5.6$  SAR is generally lower and falls in the range from 0.05 to 0.09. This difference discriminates the soils  
339 by bedrock type. Correlation between pH and SAR is statistically significant for soils on limestone ( $r = -0.632$ )

340 and sandstone ( $r=-0.648$ ), while it is statistically not significant on granite soils ( $r=-0.112$ ). Andesite and  
341 quartzite soils cannot be discussed in detail due to the limited number of samples. However, both of these soils  
342 are closer to granite soils.

343 Generally, the indices CR, MCR, and SAR used in our study appear promising for the erodibility assessment of  
344 the beech forest soils. If analyzed together with the content of major and minor elements in soils, subtle  
345 differences in soil properties on different bedrock are accentuated (Figure 2). Limestone soils are richest in the  
346 content of  $C_{org}$ , have the highest pH, the highest content of both available and total Ca and Mg, the highest  
347 clay content, and therefore are the least prone to erosion. On the contrary, granitic soils are all grouped around  
348 high erosion indices, together with sand content, indicating the greatest potential sensitivity to environmental  
349 change, i.e. through increased forest disturbance events. Sandstone soils have the highest content of Si, Al, K  
350 and Na as a consequence of the usual mineralogical composition of sandstone being dominated by quartz and  
351 feldspars.

352 The effect of changes in physical and chemical properties

353 An increase in soil erodibility in deeper layers points to higher erosion potential of the upper soil layer if the  
354 soil's physical and chemical properties are altered due to environmental change. Despite the well-established  
355 fact that soil properties change with an increase in soil depth, there is still not enough information regarding  
356 how these changes affect the potential erodibility of forest soils. The stability and content of soil organic carbon  
357 vary greatly among soil horizons as a consequence of a change in soil chemical and physical properties  
358 (Rumpel and Kögel-Knabner 2011). Conforti et al. (2016) indicated that most studies of carbon storage are  
359 focused on the topsoil horizon and neglect the deeper mineral horizons that are important for the storage of  
360 total soil organic carbon, especially in forest ecosystems. Under changing environmental conditions, which  
361 might accelerate soil erosion of the topsoil horizon, the characteristics of the subsurface soil horizons will  
362 determine the rate of further soil processes. One of the most prominent consequences of such changes in the  
363 vertical and lateral water movement through soil and nutrient leaching (Johnson 1994). Pennock and van  
364 Kessel (1997) studying medium-term plots (6 to 20 years) after clear-cutting observed higher losses of soil  
365 organic carbon, nitrogen, exchangeable Ca and Mg, soluble phosphorus, base saturation in the soil surface 0 -  
366 15 cm, while these losses were much smaller in the 15-45 cm soil horizon. However, contrasting results can be



367 found in the literature. According to a meta-analysis of forest soil carbon, James and Harrison (2016) found  
368 that the response of the soil organic carbon to harvesting in forest soils varies with depth. These authors report  
369 that the highest losses were reported to occur in the O soil horizon, followed by significant losses in the deep  
370 soils (from 60 to 100 cm).

371 Based on our results it is still not possible to propose a new erosion index that would be suitable for forest soils  
372 and it remains the task for further studies based on empirical measurements of soil transport and loss.

373 However, this index should include textural characteristics, the content of organic matter, and the content of  
374 major elements.

### 375 Conclusions

376 Under current conditions, forest soils in European mountain regions are generally not threatened by soil  
377 erosion. However, the potential of forest soils to resist erosion may dramatically change under climate change  
378 and forest management alterations. To assess susceptibility to erosion of pure beech forest soils, four  
379 erodibility indices, derived from textural (clay ratio, modified clay ratio) and geochemical (sodium adsorption  
380 ratio and oxides ratio) characteristics, were tested. The selected indices proved to be useful for the  
381 assessment of forest soil erodibility. However, none of them was sufficient when used alone, but instead, they  
382 should be combined to increase the reliability of the assessment of soil erodibility. The erodibility index most  
383 suitable for forest soils should combine textural, geochemical properties, and content of organic matter.  
384 However, based on existing indices, our results indicated that all soils under study would not erode easily  
385 under current conditions.

386  
387 We show that soil texture, the content of organic carbon, pH value, electrical conductivity, and total water-  
388 soluble cations, as components of erodibility indices, clearly differentiated forest soils by the type of bedrock,  
389 and were proved to be explanatory variables. In conditions of climate or land-use change, soil erosion can be  
390 expected to occur and erodibility would decrease in the following sequence: granitic rocks > andesite >  
391 sandstone > quartzite > limestone.

392 Environmental change will likely cause erosion of the topsoil horizon, so the characteristics of the subsurface

393 soil horizons will determine the rate of further processes. The sequence-based on the bedrock type mentioned  
394 above would be accentuated in this case. Deeper soil horizons on granitic rocks are more erodible than  
395 surface horizons, while soils on limestone do not differ in erodibility with depth.

#### 396 Acknowledgments

397 This study is a result of the COST action CA15226 Climate Smart Forestry in Mountain Regions-CLIMO.  
398 Soil analyses were financially supported by the Ministry of Education, Science and Technological Development  
399 of the Republic of Serbia Project ON176006, III43009 and Grant 453 No. 451-03-68/2020-14/200026). Michal  
400 Bosela was supported by the Slovak Research and Development Agency via projects No. APVV-15-0265 and  
401 APVV-19-0183. Tzvetan Zlatanov was supported by the Bulgarian National Science Fund (BNSF) via project  
402 No. DCOST 01/3/19.10.2018. Emira Hukić was supported by the Ministry of civil affairs of B&He project No.  
403 10-02-2-1769/20-36.

#### 404 References

- 405 Amelung, W., Blume, H., Fleige, H., Horn, R., Kandeler, E., Kögel-Knabner, I., Kretzschmar, R., Stahr, K. and  
406 Wilke., B. 2018. *Physikalische Eigenschaften und Prozesse*, Scheffer/Schachtschabel Lehrbuch der  
407 *Bodenkunde*, SpringerLink : Bücher Springer eBook Collection, 213-34, doi:10.1007/978-3-662-55871-3.
- 408 Bennett, H. H. 1926. Some comparisons of the properties of humid-tropical and humid-temperate American  
409 soils; with special reference to indicated relations between chemical composition and physical properties. *Soil*  
410 *Sci.* **21**(5):349–376.
- 411 Borrelli, P., Panagos, P., Langhammer, J., Apostol, B. and Schütt, B. 2016. Assessment of the cover changes  
412 and the soil loss potential in European forestland: first approach to derive indicators to capture the ecological  
413 impacts on soil-related forest ecosystems. *Ecol Indic* **60**:1208–1220. doi.org/10.1016/J.ECOLI ND.2015.08.053
- 414 Bouyoucos, G. J. 1935. The clay ratio as a criterion of susceptibility of soils to erosion. *J. Am. Soc. Agron.*,  
415 **27**(9):738. doi:10.2134/agronj1935.0002196200270009000
- 416 Bowditch, E., Santopuoli, G., Binder, F., del Rio, M., La Porta, N., Kluvankova, T., Lesinski, J., Motta, R., Pach,

- 417 M., Panzacchi, P., Pretzsch, H., Temperli, Ch., Tonon, G., Smith, M., Velikova, V., Weatherall, A. and Tognetti,  
418 R. 2020. What is Climate Smart Forestry? A definition from a multinational collaborative process focused on  
419 mountain regions of Europe. *Ecosyst. Serv.* **43**:101-113. doi.org/10.1016/j.ecoser.2020.101113
- 420 Bryan, R.B. 1968. The development, use and efficiency of indices of soil erodibility. *Geoderma* **478**(2):5–26.
- 421 Bünemann, E.K., Keller, B., Hoop, D., Jud, K., Boivin, P. and Frossard, E. 2013. Increased availability of  
422 phosphorus after drying and rewetting of a grassland soil: Processes and plant use. *Plant Soil*. **370**(1-2):511-  
423 526. doi.org/10.1007/s11104-013-1651-y
- 424 Ceccherini, G., Duveiller, G., Grassi, G., Lemoine, G., Avitabile, V., Pilli, R. and Cescatti, A. 2020. Abrupt  
425 increase in harvested forest area over Europe after 2015. *Nature* **583**(7814):72. DOI: 48410.1038/s41586-020-  
426 2438-y
- 427 Christensen, J.H., K. Krishna Kumar, E. Aldrian, S.-I. An, I.F.A. Cavalcanti, M. de Castro, W. Dong, P.  
428 Goswami, A. Hall, J.K. Kanyanga, A. Kitoh, J. Kossin, N.-C. Lau, J. Renwick, D.B. Stephenson, S.-P. Xie and  
429 T. Zhou, 2013: Climate Phenomena and their Relevance for Future Regional Climate Change. In: *Climate*  
430 *Change 2013: The Physical Science Basis. Contribution of Working Group I to the Fifth Assessment Report of*  
431 *the Intergovernmental Panel on Climate Change* [Stocker, T.F., Qin, D., Plattner, G.-K., Tignor, M., Allen, S.K.,  
432 Boschung, J., Nauels, A., Xia, Y., Bex, V. and Midgley, P.M. (eds.)]. Cambridge University Press, Cambridge,  
433 United Kingdom and New York, NY, USA.
- 434 Conforti, M., Lucà, F., Scarciglia, F., Matteucci, G. and Buttafuoco, G. 2016. Soil carbon stock in relation to soil  
435 properties and landscape position in a forest ecosystem of southern Italy (Calabria region), *Catena*, **144**: 23-  
436 33, doi.org/10.1016/j.catena.2016.04.023.
- 437 Cools, N. and De Vos, B. 2016. Part X: Sampling and Analysis of Soil. In: *UNECE ICP Forests Programme*  
438 *Coordinating Centre (ed.): Manual on methods and criteria for harmonized sampling, assessment, monitoring*  
439 *and analysis of the effects of air pollution on forests*. Thünen Institute of Forest Ecosystems, Eberswalde,  
440 Germany, 29 p + Annex. [<http://www.icp-forests.org/manual.htm>]
- 441 Cornes, R.C., Van der Schrier, G., Van den Besselaar, E.J.M., Jones, P.D. 2018. An Ensemble Version of the

- 442 E-OBS Temperature and Precipitation Datasets, *J. Geophys. Res. Atmos.*, **123**(17):9391-9409.  
443 doi:10.1029/2017JD028200
- 444 Dane, J.H. and Topp, G.C. 2002. *Methods of Soil Analysis, Part 4, Physical Methods*. Soil Science Society of  
445 America Book Series, **5**, Soil Science Society of America, Madison, 1692 p.
- 446 EUFORGEN <http://www.euforgen.org/species/fagus-sylvatica/>
- 447 Faulkner, H., Alexander, R. and Wilson, B.R. 2003. Changes to the dispersive characteristics of soil along an  
448 evolutionary slope sequence in the Vera badlands, southeast Spain: implications for site stabilization. *Catena*,  
449 **50**:243-254.
- 450 Faulkner, H., Spivey, D. and Alexander, R. 2000. The role of some site geochemical processes in the  
451 development and stabilization of three badland sites in Almeria, Southern Spain. *Geomorphology*, **35**:87-99.
- 452 Faulkner, H., Wilson, B.R., Solman, K. and Alexander, R. 2001. Comparison of three cation extraction methods  
453 and their use in determination of sodium adsorption ratios of some sodic soils. *Commun. Soil. Sci. Plant. Anal.*,  
454 **32**(11-12):1765-1777.
- 455 Gerbersdorf, S.U., Jancke, T. and Westrich, B. 2007. Sediment properties for assessing the erosion risk of  
456 contaminated riverine sites. *J. Soils Sediments*, **7**(1):25-35.
- 457 Grabowski, R.C., Droppo, I.G. and Wharton, G. 2010. Estimation of critical shear stress from cohesive strength  
458 meter-derived erosion thresholds. *Limnology and Oceanography-Methods*, **8**(499):678-685.
- 459 Grabowski, R.C., Droppob, I.G. and Whartona, G. 2011. Erodibility of cohesive sediment: the importance of  
460 sediment properties. *Earth Science Reviews*, **105**(3-4):101-120. [dx.doi.org/10.1016/j.earscirev.2011.01.008](https://doi.org/10.1016/j.earscirev.2011.01.008)
- 461 Haas, J., Schack-Kirchner, H. and Lang, F. 2020. Modeling soil erosion after mechanized logging operations  
462 on steep terrain in the Northern Black Forest, Germany. *Eur. J. For. Res.*, **139**:549–565.  
463 doi.org/10.1007/s10342-020-01269-5.
- 464 Hahm, W. J., Riebe, C. S., Lukens, C. E. and Araki, S. 2014. Bedrock composition regulates mountain

- 465 ecosystems and landscape evolution. *Proc. Natl Acad. Sci. USA* **111**:3338–3343.
- 466 Hartmann, D.L., Klein Tank, A.M.G., Rusticucci, M., Alexander, L.V., Brönnimann, S., Charabi, Y., Dentener,  
467 F.J., Dlugokencky, E.J., Easterling, D.R., Kaplan, A., Soden, B.J., Thorne, P.W., Wild, M. and Zhai, P.M. 2013.  
468 Observations: Atmosphere and Surface. In: *Climate Change 2013: The Physical Science Basis. Contribution of*  
469 *Working Group I to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change* [Stocker,  
470 T.F., Qin, D., Plattner, G.-K., Tignor, M., Allen, S.K., Boschung, J., Nauels, A., Xia, Y., Bex, V., Midgley, P.M.  
471 (eds.)]. Cambridge University Press, Cambridge, United Kingdom and New York, NY, USA.
- 472 IUSS Working Group WRB. 2015. World Reference Base for Soil Resources 2014, update 2015 International  
473 soil classification system for naming soils and creating legends for soil maps. *World Soil Resources Reports*  
474 106. FAO, Rome.
- 475 James, J. and Harrison, R. 2016. The effect of harvest on forest soil carbon: a meta-analysis. *Forest*,  
476 **7**(308):187-208.
- 477 Jiang, Z., Liu, H., Wang, H., Peng, J., Meersmans, J., Green, S.M., Quine, T.A., Wu, X. and Song, Z. 2020.  
478 Bedrock geochemistry influences vegetation growth by regulating the regolith water holding capacity. *Nat.*  
479 *Commun.*, **11**, 2392. doi.org/10.1038/s41467-020-16156-1
- 480 Johnson, D.W. 1994. Reasons for concern over impacts of harvesting. *In Impacts of Forest Harvesting on*  
481 *Long-term Site Productivity. Edited by Dyck, W.J., Cole, D.W. and Comerford, N.B., Chapman and Hall,*  
482 *London, Springer, Dordrecht, 371. doi.org/10.1007/978-94-011-1270-3*
- 483 Kusre, B.C., Ghosh, P. and Nath, K. 2018. Prioritization of soil conservation measures using erodibility indices  
484 as criteria in Sikkim (India). *J. Earth Syst. Sci.*, **127**(81):1-13. doi.org/10.1007/s12040-018-0981-9
- 485 Leuschner, C., Meier, I.C. and Hertel, D. 1996. On the niche breadth of *Fagus sylvatica*: soil nutrient status in  
486 50 Central European beech stands on a broad range of bedrock types. *Ann. For. Sci.* **63**:355–368. doi:  
487 10.1051/forest:2006016
- 488 Liu, M. and Han, G. 2020. Assessing soil degradation under land-use change: insight from soil erosion and soil

- 489 aggregate stability in a small karst catchment in southwest China. PeerJ, **8**:e8908. DOI 10.7717/peerj.8908
- 490 Milodowski, D.T., Mudd, S.M. and Mitchard, E.T.A. 2015 a. Erosion rates as a potential bottom-up control of  
491 forest structural characteristics in the Sierra Nevada Mountains. Ecology, **96**:31–38.
- 492 Milodowski, D.T., Mudd, S.M. and Mitchard, E.T.A. 2015 b. Topographic roughness as a signature of the  
493 emergence of bedrock in eroding landscapes, Earth Surf. Dynam., **3**:483–499, doi.org/10.5194/esurf-3-483-  
494 2015.
- 495 Minitab, LLC (2021). "Chapter Name," Getting Started with Minitab (URL: [www.minitab.com](http://www.minitab.com))
- 496 Naidu, R., Sumner, M.E. and Rengasamy, P. 1995. Australian Sodic Soils: Distribution, Properties and  
497 Management. CSIRO Publications, Melbourne, 351.
- 498 Parlak, M. 2009. Study on Splash Erosion with Interaction of Different Kinetic Energy Flux and Soil Texture. J.  
499 Agric. Sci., **15**(4):341-347.
- 500 Pennock, D.J. and van Kessel., C. 1997. Clear-cut forest harvest impacts on soil quality indicators in the  
501 mixedwood forest of Saskatchewan, Canada. Geoderma, **75**:13-32. doi.org/10.1016/S0016-7061(96)00075-4
- 502 QGIS Development Team, 2009. QGIS Geographic Information System. Open Source Geospatial Foundation.  
503 URL <http://qgis.org>
- 504 Reichstein, M., Bahn, M., Ciais, P., Frank, D., Mahecha, M. D., Seneviratne, S. I. and Wattenbach, M. 2013.  
505 Climate extremes and the carbon cycle. Nature, **500**:287–295. doi.org/10.1038/nature
- 506 Rengasamy, P., Greene, R.S.B., Ford, G.W. and Mehanni, A.H. 1984. Identification of dispersive behavior and  
507 the management of red-brown earths. Aust. J. Soil Res. **22**(541):413–431.
- 508 Robinson, D.O. and Page, J.B. 1950. Soil aggregate stability. Soil Sci. Soc. Am. Proc., **15**:25-29.  
509 doi.org/10.2136/sssaj1951.036159950015000C0005x
- 510 Rumpel, C. and Kögel-Knabner, I. 2011. Deep soil organic matter – a key but poorly understood component of  
511 terrestrial C cycle. Plant Soil, **338**:143-158. doi:10.1007/s11104-010-0391-5

- 512 Sulman, B.N., Phillips, R.P., Oishi, A.C., Shevliakova, E. and Pacala, S.W. 2014. Microbe-driven turnover  
513 offsets mineral-mediated storage of soil carbon under elevated CO<sub>2</sub>. *Nat. Clim. Change*, **4**:1099–1102.  
514 doi.org/10.1038/NCLIMATE2436
- 515 Wang, B., Zheng, F., Römkensc, M.J.M., Darboux, F. 2013. Soil erodibility for water erosion: A perspective  
516 and Chinese experiences. *Geomorphology* **187**(1):1-10 doi.org/10.1016/j.geomorph.2013.01.018
- 517 Wang, B., Zheng, F.L. and Guan, Y.H. 2016. Improved USLE- K factor prediction: a case study on water  
518 erosion areas in China. *International Soil & Water Conservation Research*, **4**:168–176. doi:  
519 10.1016/j.iswcr.2016.08.003.
- 520 Wischmeier, W.H. and Mannering, J.V. 1969. Relation of soil properties to its erodibility. Division S-6 – Soil and  
521 Water Management and Conservation. *Soil Sci. Soc. Am. J.* **56**:1560–1565.
- 522 World Topo Map URL:  
523 [https://server.arcgisonline.com/ArcGIS/rest/services/World\\_Topo\\_Map/MapServer/tile/{z}/{y}/{x}](https://server.arcgisonline.com/ArcGIS/rest/services/World_Topo_Map/MapServer/tile/{z}/{y}/{x}) (in July 2019).  
524

525 Table 1. COST Action CLIMO network beech research plots from which soils were tested

No	Country abb.	Latitude	Longitude	MAT (1961-1990) °C	MAP (1961-1990) mm	Exposition	Slope °	Altitude m a.s.l.	Bedrock	Soil type (WRB)
1	BA	43.70694444N	18.26222222E	6.7	1085	N -NW	14	1290	Limestone	Calcic Cambisol
2	BA	44.64408611N	16.66843333E	10.5	1060	E-NE	4	524	Limestone	Calcic Cambisol
3	BG	42.77916667N	23.88111111E	6.3	648	W-NW	25	1350	Sandstone	Cambisol
4	CZ	49.28516667N	16.73927778E	7.8	525	E	2.7	490	Limestone	Leptosol
5	CZ	49.28475000N	16.74008333E	7.8	525	S	4.1	485	Limestone	Leptosol
6	CZ	49.03563889N	18.01875000E	7.1	520	0	0	415	Sandstone/	Cambisol 'modal'
7	CZ	49.02344444N	18.02519444E	7.1	550	0	0	620	Sandstone/	Cambisol 'modal'
8	GE	49.06274444N	13.27144444E	6.7	1157	SW	10	720	Granite	Cambisol
9	IT	46.11888889N	12.42972222E	7.4	1749	NE	5	1090	Limestone	Luvisols
10	PL	49.43298333N	20.90310000E	5.8	929	SW	20	830	Sandstone	Cambisol
11	PL	49.62243056N	18.91460278E	7.1	1085	SW	22	520	Sandstone	Cambisol
12	RO	45.53811111N	25.91673889E	4.0	840	NE	25	1277	Quartzite	Eutric Cambisol
13	RO	45.49583333N	25.18777778E	3.4	915	NV	20	1461	Limestone	Eutric Cambisol
14	SRB	43.40625278N	21.37824722E	9.0	688	E	20	695	Granites	Cambisol dystric
15	SK	48.67796667N	19.47016667E	4.3	1004	N	10	1180	Andesite	Andic Cambisol
16	SL	46.35972222N	15.24805556E	6.9	1100	NW	32	600	Dolomite	Leptosol
17	SL	46.26083333N	15.32194444E	6.9	1100	NW	26	1070	Dolomite	Leptosol
18	SP	42.20138889N	2.721944444W	8.3	630	N	30	1430	Granite	Umbrisol
19	SP	42.20083333N	2.718611111W	8.4	630	N	23	1390	Granite	Umbrisol
20	SP	41.77555556N	2.456666667E	10.2	954	S	18	1186	Granite	Umbrisol

526 Note: BA- Bosnia and Herzegovina, BG-Bulgaria, CZ-Czech Republic, GE-Germany, IT-Italy, PL-  
527 Poland, RO-Romania, SRB-Serbia, SK-Slovakia, SI-Slovenia, and SP-Spain; MAT - mean annual  
528 temperature (°C); MAP - mean annual precipitation, Soil type is given according to WRB  
529 World reference base (IUSS Working Group WRB, 2015). Climate data are taken from:  
530 Cornes et al. (2018)  
531



532

533

Table 2. Referential values of oxides ratio ( $\sum\text{SiO}_2/\text{Fe}_2\text{O}_3+\text{Al}_2\text{O}_3$ ), sodium adsorption ratio (SAR), clay ratio

534

(CR), and modified clay ratio

Indicator	Most applicable soil texture	Benchmark values	Reference
Oxides ( $E_i = \sum\text{SiO}_2/\text{Fe}_2\text{O}_3+\text{Al}_2\text{O}_3$ )	Clay	Range 0.0-4.0 <1 – least resistant to erosion ~2 – medium resistant to erosion ~4 – very resistant to erosion	Bennett 1926
SAR	Loam	Range 0.1 – 3.0	Rengasamy 1984
	Sandy loam	Range 0.9 – 3.0	Rengasamy 1984
	Clay	Range 3.7 – 6.8	Rengasamy 1984
CR	All soil textures	Range 0.52-11.2 < - 1 very resistant to erosion > - 3 poorly resistant to erosion	Bouyoucos 1935
	Sandy	10.9-11.6	Bouyoucos 1935
	Loam	3.3	Bouyoucos 1935
	All soil textures	Range 3.28-11.00- >6.9 – proneness to erosion	Kusre et al. 2018

535

536

537 Table 3. Pearson correlation matrix showing relationships between soil erodibility indices (CR-clay ratio, MCR-  
 538 modified clay ratio, SAR-sodium adsorption ratio, oxides ratio-  $\sum\text{SiO}_2/(\text{Al}_2\text{O}_3 + \text{Fe}_2\text{O}_3)$ ), and physical and  
 539 chemical properties (C<sub>org</sub>-soil organic C in %, pH values, EC-electric conductivity, Ca<sup>++</sup> and Na<sup>+</sup> ion  
 540 concentration, TWSC-total water-soluble cations, sand content in %).

	C <sub>org</sub>	pH	EC (μS)	Ca <sup>++</sup>	Na <sup>+</sup>	TWSC	sand	MCR	CR	SAR	Oxides
C <sub>org</sub>	1	0.287*	0.686**	0.515**	0.387**	0.475**	-0.275*	-0.232	-0.054	-0.316*	-0.159
pH		1	.0501**	0.660**	-0.033	0.614**	-0.315*	-0.301*	-0.247	-0.659**	-0.221
EC			1	0.754**	0.237	0.717**	-0.366**	-0.345**	-0.201	0.634**	-0.131
Ca ion				1	0.293*	0.952**	-0.589**	-0.503**	-0.375**	-0.788**	-0.11
Na ion					1	0.328*	0.003	0.127	0.196	0.281*	-0.198
TWSC						1	-0.647**	-0.536**	-0.401**	-0.751**	-0.084
sand							1	0.924**	0.865**	0.577**	-0.227
MCR								1	0.995**	0.583**	-0.354**
CR									1	0.490**	-0.404**
SAR										1	-0.021
Oxides											1

541 \*\* Correlation is significant at the 0.01 level (2-tailed); \* Correlation is significant at the 0.05 level

542 (2-tailed). N=76 data from all bedrock groups (layer 0-10cm, 10-20cm, 20-40cm, 40-80cm) were included.

543

544

545 Figure 1. Soil sampling sites across pure beech forests in Europe. Numbers refer to site location given in Table

546 1. We used World Topo Map URL:

547 [https://server.arcgisonline.com/ArcGIS/rest/services/World\\_Topo\\_Map/MapServer/tile/{z}/{y}/{x}](https://server.arcgisonline.com/ArcGIS/rest/services/World_Topo_Map/MapServer/tile/{z}/{y}/{x}) (in July 2019)

548 we imported coordinates representing the sites from Table 1 on World Topo Map in QGIS 2009. Emira Hukić

549 created the map.

551 Figure 2. Principal component analysis for 75 pure beech forest soils on 5 bedrock types. CR-clay ratio, MCR-

552 modified clay ratio, SAR-sodium adsorption ratio, EC-electrical conductivity Ca av, Mg av, Na av, K av, Fe av,

553 Mn av, where av denotes available (water-soluble) cations. Temperature, precipitation, altitude, slope data are

554 given in Table 1.

556 Figure 3. Textural characteristics of beech forest soils.

558 Figure 4. Mean values ( $\pm$ SE) of soil erodibility indices (CR-clay ratio, MCR-modified clay ratio, SAR-sodium

559 adsorption ratio, oxides ratio- $\text{SiO}_2/(\text{Al}_2\text{O}_3 + \text{Fe}_2\text{O}_3)$ ) and soil factors (proportion of sand%-content of sand,

560  $C_{\text{org}}(\%)$ -the content of soil organic carbon, pH value, TWSC-water-soluble cation capacity) concerning soil

561 depth (1:0-10cm, 2:10-20cm, 3:20-40cm, 4:40-80cm) and different bedrock material (G-granites; A-andesite;

562 S-sandstone; Q-quartzite and L-limestone and dolostone). Gray marked surface represents all average values.



Figure 1. Soil sampling sites across pure beech forests in Europe. Numbers refer to site location given in Table 1. We used World Topo Map URL: [https://server.arcgisonline.com/ArcGIS/rest/services/World\\_Topo\\_Map/MapServer/tile/{z}/{y}/{x}](https://server.arcgisonline.com/ArcGIS/rest/services/World_Topo_Map/MapServer/tile/{z}/{y}/{x}) (in July 2019) we imported coordinates representing the sites from Table 1 on World Topo Map in QGIS 2009. Emira Hukić created the map.

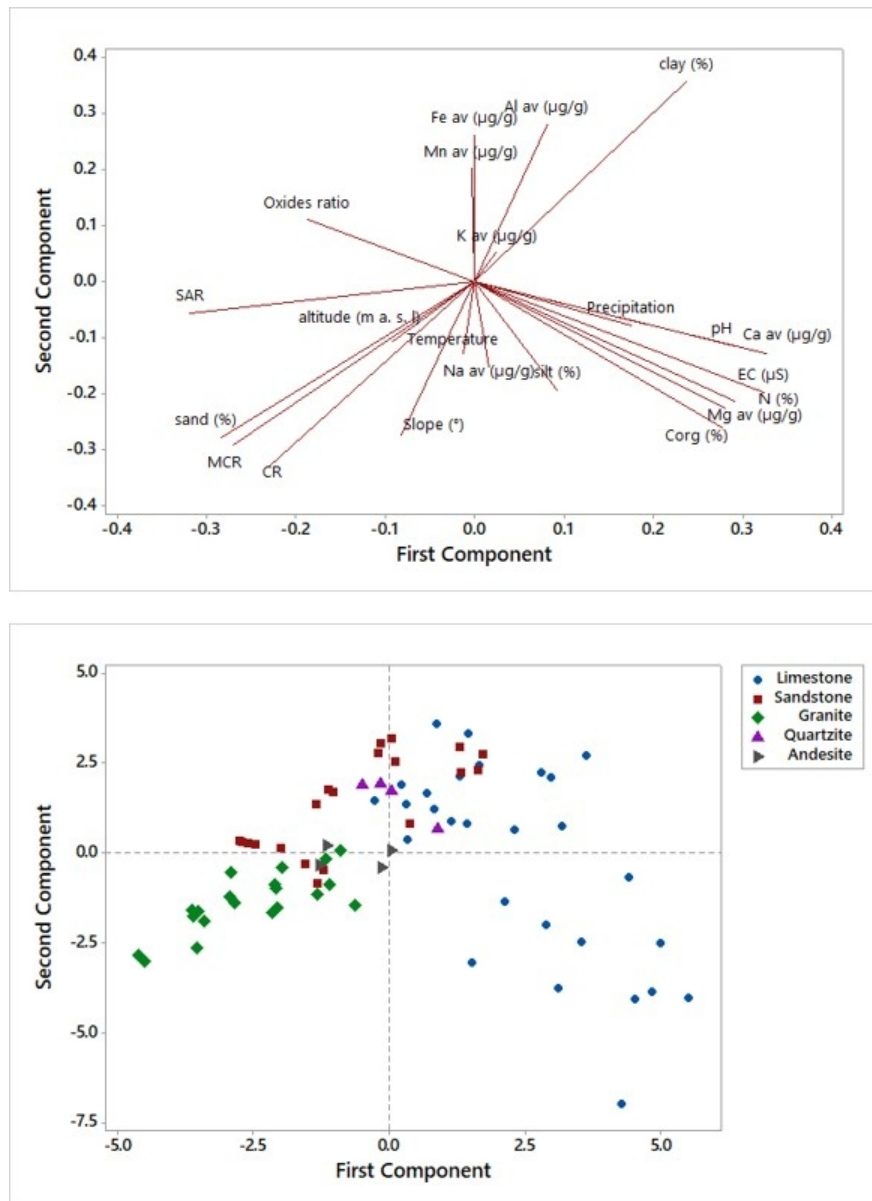


Figure 2. Principal component analysis for 75 pure beech forest soils on 5 bedrock types. CR-clay ratio, MCR-modified clay ratio, SAR-sodium adsorption ratio, EC-electrical conductivity Ca av, Mg av, Na av, K av, Fe av, Mn av, where av denotes available (water-soluble) cations. Temperature, precipitation, altitude, slope data are given in Table 1.

98x135mm (150 x 150 DPI)

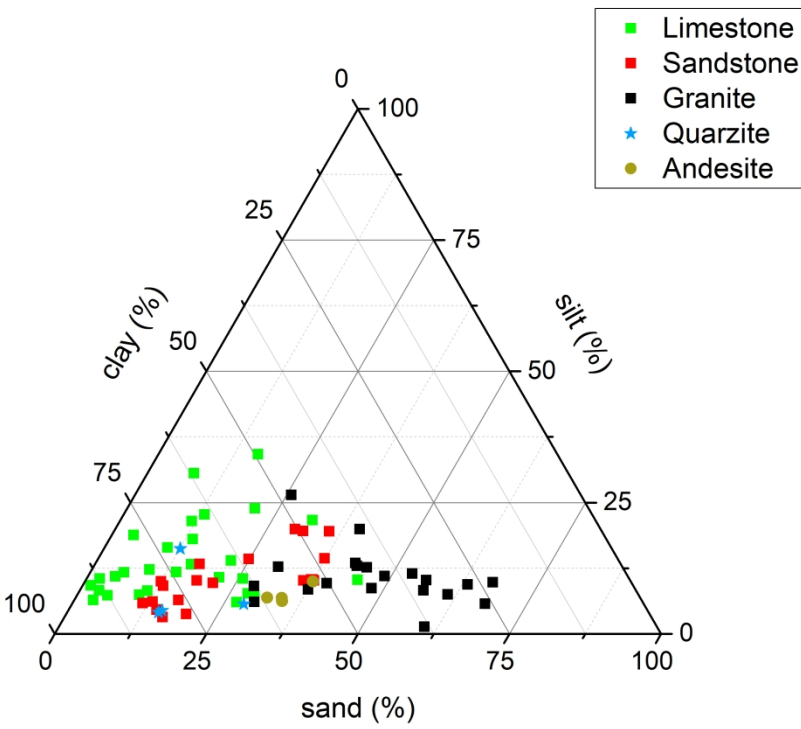


Figure 3. Textural characteristics of beech forest soils.

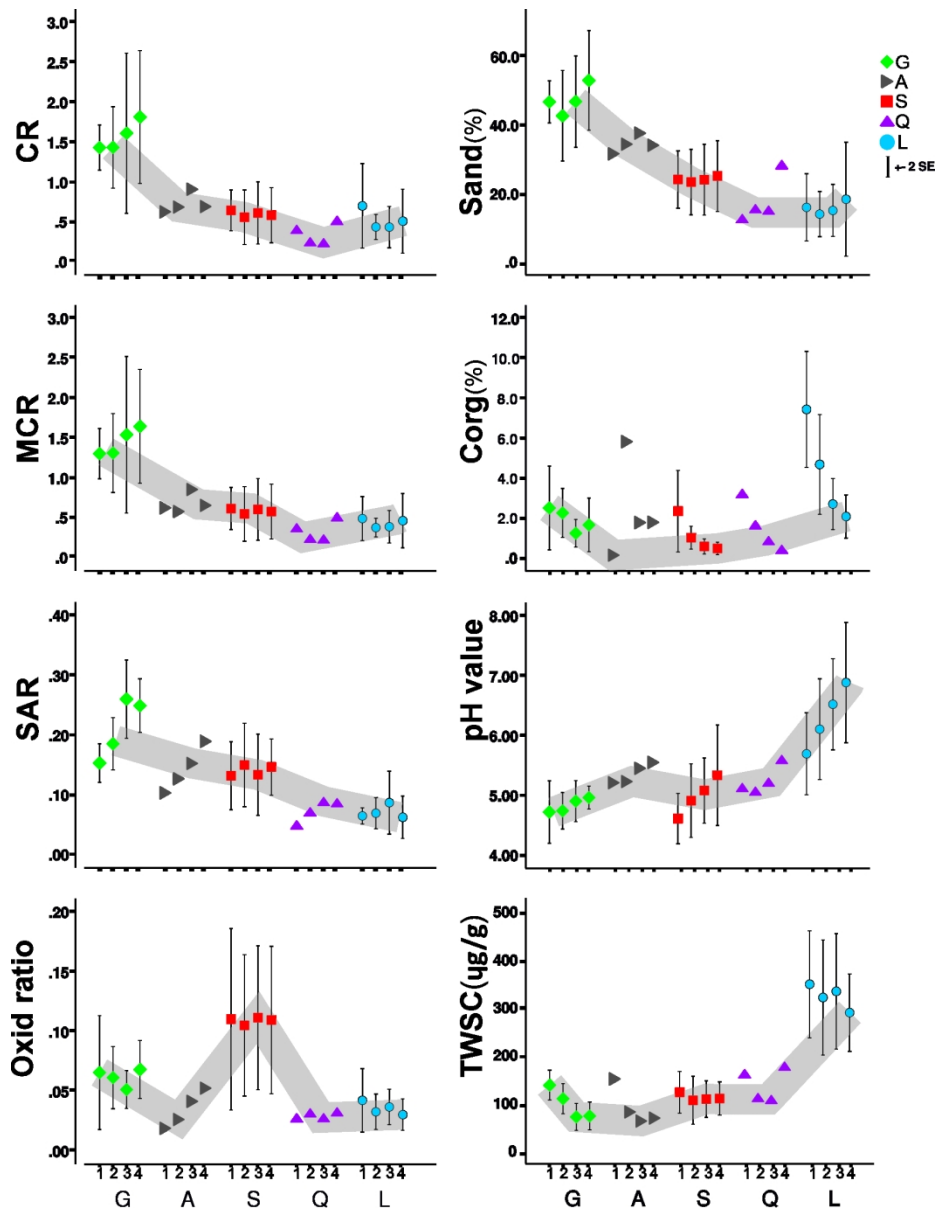


Figure 4. Mean values ( $\pm SE$ ) of soil erodibility indices (CR-clay ratio, MCR-modified clay ratio, SAR-sodium adsorption ratio, oxides ratio-SiO<sub>2</sub>/(Al<sub>2</sub>O<sub>3</sub> + Fe<sub>2</sub>O<sub>3</sub>)) and soil factors (proportion of sand%-content of sand, Corg(%)-the content of soil organic carbon, pH value, TWSC-water-soluble cation capacity) concerning soil depth (1:0-10cm, 2:10-20cm, 3:20-40cm, 4:40-80cm) and different bedrock material (G-granites; A-andesite; S-sandstone; Q-quartzite and L-limestone and dolostone). Gray marked surface represents all average values.